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Analysis of the potential for negative CO₂ emission mine sites
through bacteria-mediated carbon mineralisation:
Evidence from Australia

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Abstract

Carbon mineralisation has the potential to sequester and safely store large amounts of carbon dioxide (CO₂). However, this technology has not seen the same level of development as geological storage in sedimentary reservoirs. This is primarily because early work on carbonation of minerals focused on high temperature and pressure process routes, which are cost-prohibitive to accelerate the reaction from geological to industrial time scales. Bacteria-mediated carbon mineralisation of mine tailings [1] overcomes this impediment because it occurs at ambient temperature and standard pressure and the rock is already crushed. With abundant mafic and ultramafic rock, a highly developed mining sector, and a warm climate that promotes fast carbonation reactions, Australia presents an optimal region to develop this technology. We assess the mineralisation potential of Australia’s existing and future nickel mine tailings and estimate that this carbon sink could sequester up to 2,171 Megatons (Mt) of CO₂.

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1. Introduction

Over the last several decades, there has been growing concern that the increased concentration of carbon dioxide (\(\text{CO}_2\)) in the atmosphere may have detrimental effects on the environment and the economy [2,3]. In 1958, geochemist Charles David Keeling began to record measurements of the carbon dioxide concentration at the Mauna Loa Observatory, Hawaii. Yearly average concentration levels on the ‘Keeling curve’ have since risen sharply from 315 parts per million (ppm) in 1958 to averaging 402.24 ppm in August 2016 [4]. The level of 400 ppm has been considered to be a critical barrier for taking action to mitigate anthropogenic climate change. In 2015, the 21st yearly session of the Conference of the Parties (COP21) in Paris produced an agreement to reduce greenhouse gas emissions and keep the global average increase in surface temperature from climbing over 2°C, as compared to pre-industrial levels. This agreement was ratified by more than 190 countries [5].

Prior to COP21, participating countries were invited to initiate or intensify the domestic preparation of their respective Intended Nationally Determined Contributions (INDCs). The INDCs entail each country’s specific emission reduction goals, describing how those will contribute to achieving the objectives of the then upcoming COP21 meeting in Paris [6]. The Australian Government submitted their INDCs in August 2015 with targets covering all facets of the economy and committed to reduce greenhouse gases (GHGs) by 26-28% by 2030, relative to 2005 levels [7]. Australia faces an especially challenging undertaking because the emissions reductions must not impede the expected growth in the economy of more than 2% per annum, as outlined in the ‘Pathways to Deep Decarbonisation in 2050’ report [8].

This report is part of the global Deep Decarbonization Pathways Project (DDDP), which outlines how countries can decarbonize their economies. Among the proposed pathways for Australia, one of the key technologies identified is carbon capture and storage (CCS). In said report, CCS is foremost described as the process of capturing \(\text{CO}_2\) at point sources (e.g., power plants), transporting it and finally injecting it into an appropriate rock formation that is suitable for the long-term geological (stratigraphic) storage of \(\text{CO}_2\) [8]. The Novel \(\text{CO}_2\) Capture Task Force Report [9] provides details about how \(\text{CO}_2\) in Australia could be temporarily and permanently absorbed and stored to meet the nation’s decarbonisation goals. In addition to the ‘traditional’ CCS methods, this report also mentions a role for mineral carbonation.

The Australian Novel \(\text{CO}_2\) Capture Taskforce states that passive or accelerated carbon mineralisation technology (carbon geosequestration) may ‘genuinely and permanently sequester the necessary quantities of \(\text{CO}_2\) for centuries’ [9]. Carbon mineralisation involves a naturally occurring reaction of \(\text{CO}_2\) with magnesium- and calcium-containing rocks (mafic and ultramafic rocks) to form carbonate minerals (similar to chalk) which are composed of calcium carbonate, thereby storing \(\text{CO}_2\) in a stable and benign form [9]. The energy economics to accelerate the slow rate of passive carbonation reactions present a serious obstacle [10], resulting in the methods being unequivocally dismissed in various reports [8,9] as a viable means to contribute to the decarbonisation of the Australian economy. However, a developing technology that uses naturally occurring soil bacteria (Cyanobacteria) to induce the carbon mineralisation process without the need of external energy sources, overcomes many of these obstacles [11].

Utilizing mineral waste from ultramafic mine sites for carbon mineralisation has been documented previously [12-15]. However, when estimating the \(\text{CO}_2\) mineralisation rates, generally small sample sets have been used. More recently, complementary laboratory experiments and reactive transport modeling have been deployed to investigate and quantify the carbonation rate in mine tailings [16].

Although these strategies have proven effective in laboratory-scale studies, [1] pilot projects are necessary to evaluate strategies for accelerating carbon mineralisation, which if successful and cost effective could be incorporated into tailings storage facility design with the purpose of sequestering \(\text{CO}_2\).

Power et al. (2014) discuss these strategies and propose two scenarios for pilot projects, which if implemented at the mine-scale could render some mining operations carbon-neutral. They also provide estimates of the operational costs of the proposed scenarios on a dollar per tonne of \(\text{CO}_2\) basis to assess the economic feasibility of carbon mineralisation strategies [1].

In this paper, we assess the potential of the biological carbon mineralisation process described by Power et al. [1] to sequester \(\text{CO}_2\) across Australia. The area of focus is on a subsample of Australian nickel mines, which present optimal geological preconditions as a mineral feedstock (mafic and ultramafic host rock). We first estimate the total
sequestration potential of a subset of mines based on their past production. Next, we extrapolate these results to the entire population of Australian nickel mines. After adjusting for differences in carbon mineralisation rates by ultramafic/mafic rock source, we estimate the Australia-wide carbonation potential based on existing nickel mines, though platinum group metals, chrysotile asbestos, diamond, chromite, and talc mines can also enhance carbon mineralisation. Finally, we will estimate potential CO₂ sequestration, based on past production and known Australian reserves. The key questions that are dealt with in this paper are the following:

- What is the total available mafic and ultramafic mineral feedstock from both past and future mine production?
- What is the potential of bacterially mediated carbon mineralisation to sequester CO₂ across nickel and other ultramafic/mafic mines in Australia?
- What is the yearly uptake potential of carbon dioxide through bacterially mediated carbon mineralisation?

2. Methods

2.1. Study area: mafic and ultramafic nickel mine sites

Using Geoscience Australia’s Archean, i.e., mafic and ultramafic magmatic, events map (Figure 1), the continental scale of carbon mineralization that could be realized from events producing rock units that cover 100’s of km² can be appreciated. For example, a single rock unit 20 km² by 50 m in depth would hold in excess of 10¹⁰ metric tons of ultramafic rock that could sequester 1 Gt of carbon (Table 1, Reactions 1 - 5).

Figure 1. Map of Australia’s Archean mafic and ultramafic magmatic events [18]

<table>
<thead>
<tr>
<th>Chemical reactions [17]</th>
<th></th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Silicate weathering</strong>:</td>
<td>Mg₃Si₂O₅(OH)₄ + 3 H₂SO₄ → 3 Mg²⁺ + 2 Si(OH)₄ + H₂O + 3 SO₄²⁻</td>
</tr>
<tr>
<td><strong>Carbon Mineralisation</strong>:</td>
<td>5 Mg²⁺ + 4 CO₃²⁻ + 2 OH⁻ + 5 H₂O → Mg₅(CO₃)₄(OH)₂·5H₂O</td>
</tr>
<tr>
<td><strong>Photosynthesis</strong>:</td>
<td>HCO₃⁻ + H₂O → CH₂O + O₂ + OH⁻</td>
</tr>
<tr>
<td><strong>Carbon dioxide equilibrium with water</strong>:</td>
<td>CO₂ + H₂O ⇌ HCO₃⁻ + H⁺</td>
</tr>
<tr>
<td><strong>Carbonate ion formation</strong>:</td>
<td>HCO₃⁻ + OH⁻ → CO₃²⁻ + H₂O</td>
</tr>
</tbody>
</table>

Table 1. Chemical reactions [17]

Ultramafic rock typically contains a high percentage (up to 50 %, by weight) of magnesium oxide. Ophiolite belts, which are comprised of an abundance of magnesium as serpentine [Mg₃Si₂O₅(OH)₄], olivine (Mg₂SiO₄) and brucite [Mg(OH)₂] represent the best feedstock for carbon mineralisation [19]. Mafic and ultramafic mine sites are a valuable feedstock for carbon mineralisation and there is potential to be used to offset CO₂ emissions that are generated by the mining industry. It has been documented that passive carbonation is already occurring at various mine sites (for example, Mount Keith, Australia, Clinton Creek, Canada and Diavik, Canada) [1]; however, the potential remains largely untapped.
Currently, we do not know the amount of mine tailings that are generated annually from the production of nickel, platinum group elements, chrysotile asbestos, diamond, chromite and talc. However, data for individual ore deposits is available; for instance, the nickel mine at Mount Keith in Western Australia produces approximately 11 Mt of tailings each year. The complete carbonation of the tailings could sequester approximately 4 Mt of CO₂ every year, which represents 10 times the annual emissions of Mount Keith [1]. Although passive carbonation is ongoing at the Mount Keith mine site, and accounts for roughly 11% of total GHG emissions per annum, the core potential remains untapped [20].

2.2. Focus: Nickel mine sites

Nickel mines are the area of concentration in this paper, as they present optimal feedstock (mafic and ultramafic rocks) and both active and abandoned sites can be found in large numbers across Australia. As of September 2016, the Australian Mine Atlas lists 129 mines (historic, operational, planned) as well as 207 deposits containing nickel [21]. Australia is one of the largest nickel producers in the world with a production of 246 kt and holds approximately one quarter of the world’s estimated economic nickel resources, which equated to 81 Mt in 2014 [22].

Nickel can commonly be found in two main ore types: sulfide and laterite ores. Sulfide often includes copper and/or cobalt and sometimes also precious metals such as gold, platinum or palladium (PGMs). Laterite ores are formed near Earth’s surface after extensive weathering and occur near the equator or in arid regions such as southern Africa or Western Australia [23]. In the past, most nickel production was derived from sulfide ores, due to lower cost and less complex processing requirements compared to laterite ores. In terms of resources, however, approximately 60% of Ni reserves are found in laterites whereas 40% are contained in sulfide-bearing ore [23].

2.3. Approach

The approach for estimating sequestration potential starts with an assessment of past nickel ore production across some of the largest mine sites in Australia (Figure 2).

The list has been composed with the aid of the detailed research report about the sustainability of mining in Australia [23]. The mineralogical and geological data have been assessed in detail, in order to determine the potential for biologically mediated carbon mineralisation. Total mineral feedstock values from the sample dataset were extended to include the remainder of Australia’s nickel production. It includes estimates of future production of nickel ores; both for the selected mine sites as well as the remainder of Australia’s nickel production. Since estimates of future potential are based on less precise estimates of mineralogy and geology, we have elected to select the more conservative estimates to avoid overstating the carbon mineralisation potential.

The carbon mineralisation conversion table (Table 2) translates the mineral feedstock into sequestration potential of carbon dioxide within various magnesium carbonate minerals. It lists various mafic and ultramafic minerals that yield differing feedstock-to-product ratios. For example, brucite, which is a highly reactive mineral, can sequester
one ton of CO$_2$ for every 1.33 ton of mineral feedstock if the mineral product has a CO$_2$:MgO ratio of 1:1. Contrastingly, production of artinite and pokrovskite, which have lower CO$_2$:MgO ratios, would require carbonation of 2.65 tons of brucite to sequester one ton of CO$_2$ (Table 2).

Table 2. Carbon mineralisation table [reproduced from 24]

<table>
<thead>
<tr>
<th>Mineral/ Formula</th>
<th>CO$_2$:MgO Ratio</th>
<th>Serpentine Mg$_3$Si$_2$O$_5$(OH)$_4$</th>
<th>Brucite Mg(OH)$_2$</th>
<th>Forsterite Mg$_2$SiO$_4$</th>
<th>Diopside CaMgSi$_2$O$_6$</th>
<th>Enstatite Mg$_2$SiO$_4$</th>
</tr>
</thead>
<tbody>
<tr>
<td>Magnesite, MgCO$_3$</td>
<td>1:1</td>
<td>2.10</td>
<td>1.33</td>
<td>1.60</td>
<td>2.46</td>
<td>2.28</td>
</tr>
<tr>
<td>Hydromagnesite, (Mg$_6$(CO$_3$)$_4$(OH)$_2$·4H$_2$O</td>
<td>4:5</td>
<td>2.62</td>
<td>1.66</td>
<td>2.00</td>
<td>2.73</td>
<td>2.85</td>
</tr>
<tr>
<td>Dypingite, Mg$_6$(CO$_3$)$_4$(OH)$_2$·5H$_2$O</td>
<td>4:5</td>
<td>2.62</td>
<td>1.66</td>
<td>2.00</td>
<td>2.73</td>
<td>2.85</td>
</tr>
<tr>
<td>Pokrovskite, Mg$_6$(CO$_3$)$_4$(OH)$_2$</td>
<td>1:2</td>
<td>4.20</td>
<td>2.65</td>
<td>3.20</td>
<td>3.28</td>
<td>4.56</td>
</tr>
<tr>
<td>Artinite, Mg$_6$(CO$_3$)$_4$(OH)$_2$·3H$_2$O</td>
<td>1:2</td>
<td>4.20</td>
<td>2.65</td>
<td>3.20</td>
<td>3.28</td>
<td>4.56</td>
</tr>
<tr>
<td>Nesquehonite, Mg$_6$CO$_3$·3H$_2$O</td>
<td>1:1</td>
<td>2.10</td>
<td>1.33</td>
<td>1.60</td>
<td>2.46</td>
<td>2.28</td>
</tr>
<tr>
<td>Lansfordite, MgCO$_3$·5H$_2$O</td>
<td>1:1</td>
<td>2.10</td>
<td>1.33</td>
<td>1.60</td>
<td>2.46</td>
<td>2.28</td>
</tr>
</tbody>
</table>

Combining the conversion table with the previously identified mineral feedstock for both past and future production will then give a range of the amount of CO$_2$ that could potentially be sequestered across Australia’s nickel mine sites. The last step will look into methods for absorbing the estimated amounts of CO$_2$ utilizing heap-leaching and a photosynthetic bioreactor to accelerate carbon mineralisation.

### 2.4. Ultramafic feedstock from nickel mines: estimations

The first element for determining the CO$_2$ sequestration potential is the past production of nickel at various mine sites across the country. Based on data from Mudd [23] with regards to nickel mines in Australia, key mine sites have been selected. The data from these selected mine sites have been enriched with publicly available information from the peer-reviewed literature as well as company announcements [25-40]. In a second step, the remainder ore from Australian nickel production (past and future) was estimated.

Table 3 lists nickel sulfide mines and nickel laterite mines separately. While nickel production from sulfide mines has made up the vast majority of past production, laterite mines have approximately double the reserves compared to sulfide mines. Therefore, our calculations predict that nickel from lateritic mines will make up the majority of future Australian nickel production [23].

The selected sulfide and laterite nickel mines comprise of a total of 23 mine sites, most of which are found in Western Australia. This list represents some of the largest and most important nickel mines in Australia that have adequate mineralogical and geological data available to estimate sequestration potentials. These mine sites have produced 3,812 kt of nickel and approximately 372 Mt of ore. The estimated reserves of the selected mine sites equate to 23,316 kt of nickel, which translates into 3,202 Mt of nickel ore.

The lower half of Table 3 lists past nickel production for the remainder of Australian mine sites, which totals 3,347 kt of nickel (as per 2015) and an estimated ore production of 438 Mt. Furthermore, Geoscience Australia has estimated the Economic Demonstrated Reserves (EDR) of nickel to be 19,000 kt, whereas the Subeconomic Demonstrated Reserves (SDR) are estimated to be 4,000 kt [41]. These are Economic Demonstrated Resources (EDR) measures that are established and either analytically demonstrated, or can be assumed with a reasonable certainty that extraction would be profitable under predefined investment assumptions. Subeconomic Demonstrated Resources (SDR) is comparable to EDR regarding certainty of occurrence. However, financial viability of SDR might only be reflected in the future [41]. Since the ‘Selected nickel mines’ section covers the main nickel projects in Australia, the EDR and SDR are considered to be included therein and will not be additionally listed as mining and sequestration potential. The inferred Resources (IR) are listed separately, as they present resources for which quantitative estimates are based on a broad knowledge of the geological character of the deposit. Therefore, there are either none or only few samples or measurements [42]. It is assumed for the purpose of this paper that the nickel concentration for the inferred resources is 0.5 %, which is lower than past production as concentration is declining over time [23].
Table 3. Production and reserves of mafic and ultramafic feedstock from nickel mines

<table>
<thead>
<tr>
<th></th>
<th>Past nickel production (kt nickel)</th>
<th>Ø Nickel concentration (%)</th>
<th>Past ore production (Mt ore)</th>
<th>Total Reserves (kt nickel)</th>
<th>Ø Nickel concentration (%)</th>
<th>Total Reserves (Mt ore)</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Selected nickel mines</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Nickel sulfide mines</td>
<td>3,434</td>
<td>1.02</td>
<td>336</td>
<td>8,271</td>
<td>0.67</td>
<td>1,210</td>
</tr>
<tr>
<td>Nickel laterite mines</td>
<td>378</td>
<td>1.06</td>
<td>36</td>
<td>15,189</td>
<td>0.76</td>
<td>1,992</td>
</tr>
<tr>
<td><strong>Total selected mines</strong></td>
<td>3,812</td>
<td>1.02</td>
<td>372</td>
<td>23,316</td>
<td>0.73</td>
<td>3,202</td>
</tr>
<tr>
<td><strong>Remainder of Australia</strong></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>1967 – 2007 (Mudd)</td>
<td>1,857</td>
<td>0.76</td>
<td>243</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>2008 – 2015 (OCE)</td>
<td>1,490</td>
<td>0.76</td>
<td>195</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Economic Demonstrated Resources (Included in ‘Selected nickel mines’)</td>
<td>19,000</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Subeconomic Demonstrated Resources (Included in ‘Selected nickel mines’)</td>
<td>4,000</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Inferred Resources</td>
<td>17,800</td>
<td>0.50</td>
<td>3,560</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td><strong>Total remainder of Australia</strong></td>
<td>3,347</td>
<td>438</td>
<td>17,800</td>
<td>3,560</td>
<td></td>
<td></td>
</tr>
<tr>
<td><strong>Total Australia</strong></td>
<td>7,159</td>
<td>810</td>
<td>41,116</td>
<td>6,762</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

* assumed concentration

2.5. Carbon sequestration potential

The geologic setting, production rates, reserves and mineralogy of key Australian sulfidic and laterite-hosted nickel mines (23 locations across Australia) are quite well defined. Therefore, the sequestration potential for past production as well as the reserves for these sites could be estimated using publicly available data. Where mineralogical data were not available, the information for a ‘typical’ mine type (based on the existing sulfide and laterite mine data) was applied.

The estimated mafic and ultramafic feedstock established in Table 3 will serve as a base for the calculations of carbon sequestration potential. The uptake potential is dependent on the subtype of feedstock (Table 2). In order to estimate the carbon sequestration potential from nickel mining, the mafic and ultramafic feedstock established in Table 3 were used to calculate the carbonation yields, using the conversion factors provided by Power et al., ([24]; see Table 2), which range from 1.66 tons (for brucite) to 2.85 tons (for enstatite) to fix one ton of CO\(_2\) as dypingite or hydromagnesite, which commonly form in Australian mine tailings (e.g., [16]). This project focuses on dypingite and hydromagnesite as target carbonates because they are more stable than lansfordite or nesquehonite, and their formation can be bacterially mediated [10,14,43].

For the remainder of Australia, both past production and reserves, the mineralogical setting has yet to be defined. Therefore, a mix of the aforementioned profiles (sulfide and laterite) was assumed. The ratios applied for the ‘Remainder of Australia’ section are 0.9 sulfide to 0.1 laterite (for past production) and 0.3 sulfide to 0.7 laterite (for estimated reserves of the inferred resources). This was adjusted for the fact that future production of nickel will likely stem predominantly from laterite mine sites, as there is a greater abundance of these resources to be exploited [23]. The CO\(_2\) sequestration capacity assumes complete carbonation of the mineral feedstock.

A total of 2,171 Mt of CO\(_2\) could be sequestered considering past production and future production of both the selected mine sites as well as the inferred resources. Nickel sulfide mine sites make up the largest portion of the sequestration potential from past production of the selected mine sites, with approximately 111 Mt of CO\(_2\), while lateritic mine sites hold a potential of 12 Mt of CO\(_2\). An additional 148 Mt of CO\(_2\) could be sequestered from the past production of the remainder of Australia’s nickel mines, totaling at 271 Mt of CO\(_2\). The past to future production sequestration ratio is 1:7, with 271 to 1,900 Mt of CO\(_2\).
Table 4. CO₂ sequestration potential from nickel mines feedstock

<table>
<thead>
<tr>
<th>Past ore production (Mt ore)</th>
<th>Total Reserves (Mt ore)</th>
<th>Sequestration from past production (Mt CO₂)</th>
<th>Sequestration from future production (Mt CO₂)</th>
</tr>
</thead>
<tbody>
<tr>
<td><strong>Selected sulfide and laterite nickel mines</strong></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Nickel sulfide mines</td>
<td>336</td>
<td>1,210</td>
<td>111</td>
</tr>
<tr>
<td>Nickel laterite mines</td>
<td>36</td>
<td>1,992</td>
<td>12</td>
</tr>
<tr>
<td><strong>Total selected mines</strong></td>
<td><strong>372</strong></td>
<td><strong>3,202</strong></td>
<td><strong>123</strong></td>
</tr>
<tr>
<td><strong>Remainder of Australia</strong></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>1967 – 2007 (Mudd)</td>
<td>243</td>
<td></td>
<td>82</td>
</tr>
<tr>
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<td></td>
<td>66</td>
</tr>
<tr>
<td>Inferred Resources (IR)</td>
<td></td>
<td>3,560</td>
<td></td>
</tr>
<tr>
<td><strong>Total remainder of Australia</strong></td>
<td><strong>438</strong></td>
<td><strong>3,560</strong></td>
<td><strong>148</strong></td>
</tr>
<tr>
<td>Sequestration (Mt CO₂)</td>
<td></td>
<td></td>
<td>271</td>
</tr>
<tr>
<td><strong>Total sequestration (Mt CO₂)</strong></td>
<td></td>
<td></td>
<td><strong>2,171</strong></td>
</tr>
</tbody>
</table>

2.6. Annual CO₂ uptake estimations

A direct biotechnological approach to carbon mineralisation could be devised at mine sites by combining heap leaching, which has been used for centuries to extract metals from sulfidic low-grade ore and a photosynthetic bioreactor [1,44]. Briefly, this could be done by placing the ore on a liner and inoculating with acid generating bacteria and nutrients. Sulfidic material provides ideal conditions for acid-generating microbes, such as *Acidithiobacillus spp.* that accelerate the generation of sulfuric acid for mineral dissolution and metal leaching [1,44]. For example, leaching 7500 tons of ultramafic mine tailings, could release enough magnesium to sequester approximately 2,800 tons of CO₂ per year, in a 0.12 km² photosynthetic bioreactor.

3. Results

Leading up to the COP21 in Paris as of November 2015, the Australian government had pledged to reduce its overall GHGs by 26 to 28 % in 2030 compared to 2005 levels [7]. As of 2014, Australia emitted a total of approximately 523 Mt of CO₂ equivalent gases (CO₂e), whereas 2005 emissions amounted to 519 Mt of CO₂e [45]. Therefore, in order to reach its pledge, the annual CO₂e would have to be reduced to 376 CO₂e, which represents a reduction of approximately 150 Mt (or 28 %) of CO₂e by 2030. Assuming linearity, an average additional reduction of approximately 11 Mt of CO₂e per annum would have to occur over the following 14 years in order to reach the goals stated at the COP21. In other words, a total amount of roughly 1,115 Mt of CO₂ would need to be reduced in total from 2016 to 2030.

While the Australian government will be working on a suite of solutions in order to achieve the pledged goals, bacterially mediated carbon mineralisation could be a substantial contributor. Based on the current data for Australia there is potential for sequestering more than 2,000 Mt of CO₂ utilizing past and future feedstock. If bioreactors were used to accelerate this bacteria-mediated mineralisation process to full potential by 2030, this would result in more than 10x the required reductions.

4. Discussion

Carbon mineralisation has been identified by the Australian Novel CO₂ Capture Task Force Report as being one of few technologies that have the potential to sequester and safely store large amounts of CO₂ for centuries. Australia
is an optimal region for further developing this technology due to the abundance of mafic/ultramafic rock and mineral waste that is rich in magnesium [9].

Using mine waste as a feedstock possesses several additional benefits. First, as the rocks are crushed and ground, their high reactive surface area presents an optimal material for carbon mineralisation reactions. Furthermore, they provide an incentive for mining companies to market their reduced carbon footprint. To accelerate the rate of naturally-occurring carbon mineralisation, most processes rely on very high pressure and temperature and the process costs associated with these increased reaction kinetics have rendered these processes economically unviable. Bacteria-mediated carbon mineralisation overcomes eliminates the need for much of the energy required to generate the high pressures and temperatures required to mediate the reaction.

Although this study demonstrates how bacterially-mediated carbon mineralisation of mine tailings has the potential to overcome the previous energy economics objections, this study has not demonstrated its economic viability, especially given the current price of carbon. Economists concur that carbon is currently undervalued [46]; while the ‘social cost of carbon’ is estimated at USD 200 according to Pindyck’s [46], there are dozens of different market prices in the form of national carbon floors or cap and trade systems with a proliferation of announcements in 2016 from both emerging and developed countries. Ernst and Young [47] predict that beginning in 2020, companies are likely to be ‘carbon constrained in all the major emitting and emerging countries’ and expect a global carbon market to appear. A global carbon market will remove much of the uncertainty in the benefits of the process. Future and ongoing work strives to integrate the costs associated with these processes.

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References

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